Effects of climate change and adaptation on the livestock component of mixed farming systems: A modelling study from semi-arid Zimbabwe

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ABSTRACT

Large uncertainties about the impacts of climate change and adaptation options on the livestock component of heterogeneous African farming systems hamper tailored decision making towards climate-smart agriculture. This study addressed this knowledge gap through the development and use of a dynamic modelling framework integrating climate, crop, pasture and livestock models. The framework was applied to a population of 91 farms located in semi-arid Zimbabwe to assess effects on livestock production resulting from climate change and management interventions. Climate scenarios representing relative "cool-wet", "hot-dry" and "middle" conditions by mid-century (2040–2070) for two representative concentration pathways were compared with the baseline climate. On-farm fodder resources and rangeland grass production were simulated with the crop model APSIM and the pasture model GRASP respectively. The simulated fodder availability was used in the livestock model LIVSIM to generate various production indicators including milk, offtake, mortality, manure, and net revenue. We investigated the effects of two adaptation packages targeting soil fertility management and crop diversification and quantified the sensitivity to climate change of both current and improved systems. Livestock productivity was constrained by dry-season feed gaps, which were particularly severe for crude protein and caused by the reliance on rangeland grazing and crop residues, both of low quality in the dry season. Effects on grass and stover production depended on the climate scenario and the crop, but year-to-year variation generally increased. Relative changes in livestock net revenue compared to the baseline climate varied from a 6% increase to a 43% decrease, and the proportion of farmers negatively affected varied from 20% to 100%, depending on the climate scenario. Adverse effects of climate change on average livestock production usually coincided with increased year-to-year variability and risk. Farms with larger stocking density faced more severe feed gaps and were more sensitive to climate change than less densely stocked farms. The first adaptation package resulted in increased stover production and a small increase in livestock productivity. The inclusion of grain and forage legumes with the second package increased milk productivity and net revenues more profoundly by 30%. This was attributed to the alleviation of dry-season feed gaps, which also reduced the sensitivity to climate change compared to the current system. Clearly, individual farms were affected differently by climate change and by improved farm management, illustrating that disaggregated impact assessments are needed to effectively inform decision making towards climate change adaptation.

1. Introduction

Smallholder farming systems are vulnerable to climate change and likely to be adversely affected to varying extents across sub-Saharan Africa (Naab et al., 2012; Descheemaeker et al., 2016a). Mixed crop-livestock systems are the predominant farming system throughout the semi-arid and sub-humid zones of the continent with important contributions to meat and milk production, to crop production through the provision of traction and manure, and to livelihoods for millions of rural people (Herrero et al., 2010; Tarawali et al., 2011). To improve the resilience of these farming systems, context-specific information is needed for effective decision making and for the selection and implementation of strategies towards climate-smart agriculture (Lipper et al., 2014; Thornton and Herrero, 2014). However, large

uncertainties and knowledge gaps persist about the likely effects of climate change and adaptation options, in particular for effects on the livestock component (Weindl et al., 2015) and for heterogeneous farm populations (see Descheemaeker et al. (2016a) for a review of the gaps).

Farming systems in sub-Saharan Africa are diverse, with enormous heterogeneity between households in terms of objectives, attitudes and access to natural, financial, physical, human and social capitals (Giller et al., 2011; Descheemaeker et al., 2016b). Hence, the impact of climate change and adaptation options is likely to differ between farm types (Masikati et al., 2015; Traore et al., 2017), and should not be generalized in impact assessments (Thornton et al., 2007). Integrated assessments can inform strategies towards climate-smart agriculture (Claessens et al., 2012; Antle et al., 2016), but need data produced with detailed, process-based models that allow simulating the effects of climate and adaptation options on the biophysical components of the farm (e.g. crops, livestock, soils). Yet, especially for the livestock component, methods to quantify these effects for heterogeneous African farming systems are only now being described and tested (Rodriguez et al., 2017).

Livestock are affected by climate change through changes in feed resources, including their quantity, quality and temporal and spatial distribution, changes in temperature (heat stress), changes in the availability and quality of water resources and changes in disease occurrence and pressure (Thornton et al., 2009; Godber and Wall, 2014). In mixed smallholder farming systems, feed resources include grazed biomass from rangelands, crop residues and, to a lesser extent, forages and concentrates (Valbuena et al., 2015). Each of these feed resources may be affected by climate change in different ways. Smallholders usually keep livestock for multiple purposes, including functions for which herd size matters more than individual animal productivity, such as insurance, banking, socio-cultural and crop-supporting (manure, traction) functions (Moll, 2005; Mekonnen et al., 2011). As such, excessively large herd sizes often compromise the efficiency of milk and meat production. The different functions of livestock might be affected differentially by climate change and adaptation options, but also here, little is known about these impacts.

Semi-arid Zimbabwe was chosen as a case study, as it is representative for large areas of semi-arid land in southern Africa where rainfed agricultural production is the mainstay of the rural population but increasingly under threat from climate change (Masikati et al., 2015). Southern Africa is expected to be strongly exposed to the adverse effects of climate change, with a predicted temperature increase by the end of the century of up to 3–6 °C, combined with likely less and more variable rainfall (Niang et al., 2014). The reliable crop growing days in the study area are expected to drop below 90 by the year 2050 (Jones and Thornton, 2009). In such conditions, it is expected that rainfed crop production will become increasingly risky and farmers will shift to livestock keeping. This trend is likely to be reinforced by the increasing demand for livestock products in the region (OECD/FAO, 2016). However, notwithstanding the importance of the sector, livestock has received very little attention in regional policy documents aimed at climate adaptation (van Garderen, 2011).

In this paper we present and use a modelling framework for assessing impacts on the livestock component of mixed systems in heterogeneous farm populations. We start with describing the farming system in the study area, and in particular the intake of different feed types over time. We then assess the impact of climate change and of two adaptation packages on the feedbase and on livestock production, while taking into account the uncertainty associated with different climate scenarios. In doing so, we test the hypothesis that different types of farms are affected differently by climate change and also by improved management. We further investigate whether the improved system is less sensitive to climate change than the current system.

2. Materials and methods

2.1. Study area

The rural district of Nkayi in Natural Region IV of Zimbabwe was chosen as a representative study area for semi-arid areas that cover about a third of Zimbabwe (Homann-Kee Tui et al., 2013b). The average annual rainfall in Nkayi is 650 mm with high interannual variability, and minimum and maximum mean temperatures are 15 °C and 30 °C respectively. With > 76% of the rural population below the poverty line (ZimVAC, 2013) and food self-sufficiency varying from 3 to 10 months depending on rainfall, rural households are extremely vulnerable to the adverse effects of climate change. Low and variable rainfall, combined with poor soil fertility and limited input use result in low agricultural productivity (Homann-Kee Tui et al., 2013b). Farming systems in Nkayi are mixed crop-livestock systems (Homann-Kee Tui et al., 2015a), in which crop residues are used as dry season feed, and livestock provides draft power and manure to crop production. All farmers cultivate maize (*Zea mays* L.), whereas about a third grows sorghum (*Sorghum bicolor* (L.) Moench) and pearl millet (*Pennisetum glaucum* (L.) R.Br.) and a third includes groundnut (*Arachis hypogaea* L.). Current crop yields are low, with maize yielding on average 0.7 t ha^{-1} and millet and groundnut not surpassing 0.5 t ha^{-1}. About 60% of the households keep cattle and/or goats and donkeys, but productivity is also poor (Homann-Kee Tui et al., 2015a). Communal rangelands provide the major part of the livestock feedbase (Homann-Kee Tui et al., 2013b).

2.2. Modelling framework and data

We followed the Agricultural Model Intercomparison and Improvement Project (AgMIP) Regional Integrated Assessment (RIA) approach that links climate, crop, livestock and economic data and models for assessing the effect of climate change and adaptation options on heterogeneous farm populations (AgMIP, 2015). In this paper, we focused on the livestock component, and its links with the crop and rangeland components of the farming system. In our modelling framework (Fig. 1), field-level information on crop yields, community-level information on rangeland biomass, and herd-level information on animals is integrated with farm-level information on cropland allocation and soil, crop and livestock management practices. We assessed effects of climate change and two adaptation packages on feed availability and, through the feed, on various livestock outputs for all cattle-keeping households in a farm population. The AgMIP RIA models were calibrated for current production systems and run for the current and mid-century period (2040–2070). A detailed description of the methodology for the separate models can be found in the AgMIP RIA handbook (AgMIP, 2015), Antle et al. (2015) and Masikati et al. (2015). In what follows, we provide more details on the livestock model LIVSIM and the data and models feeding into LIVSIM.

2.2.1. Household information

Village and household data were collected in 2011 from 8 villages and a total of 160 households using individual interviews and group discussions. The sampling and interview methods are described in Homann-Kee Tui et al. (2013a). Farm households were stratified into three types (the extremely poor, poor and non-poor) based on resource endowments (Masikati et al., 2015). The modelling framework was run with specific settings for each household. Household-specific information on soil types, fertilizer rates and sowing windows were used in the crop growth model APSIM to simulate grain and stover yields of the major crops (see section 2.2.3). In combination with cropland areas, stover and forage yields were used to calculate farm-level feed availability for use in LIVSIM. Community-level information on rangeland areas and number of animals allowed estimating the rangeland stocking rate, which was used in combination with the grass simula-
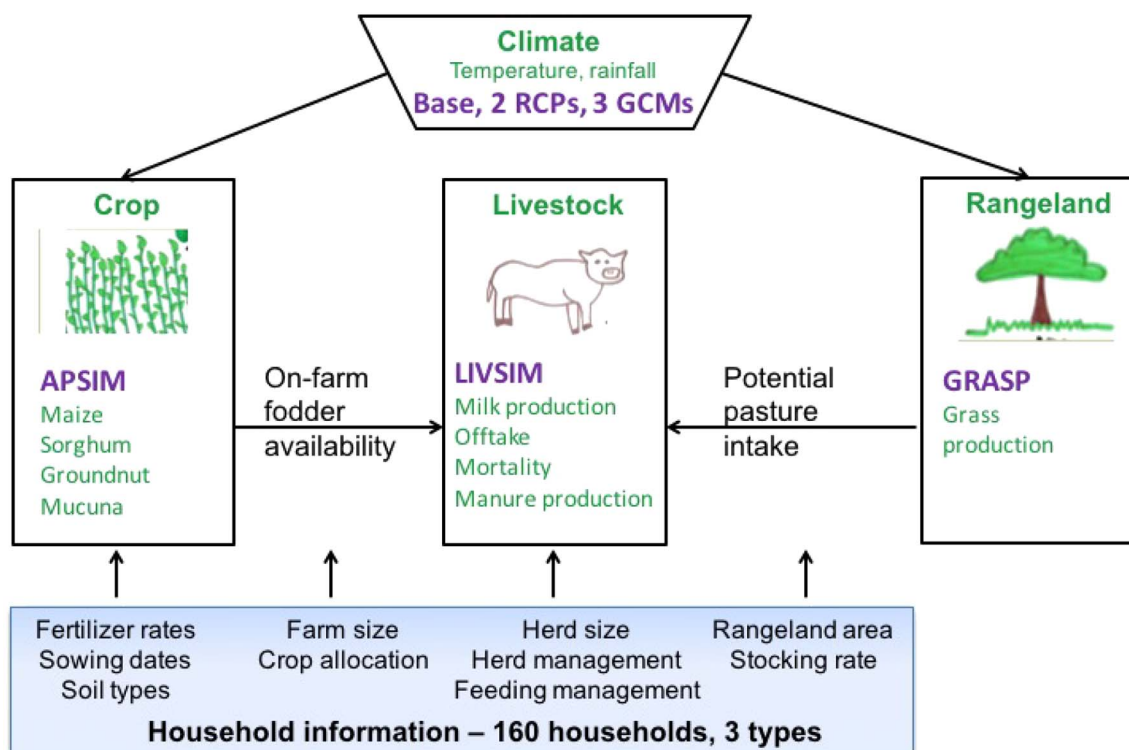
Fig. 1. Conceptual representation of the modelling framework, with indication of the simulated components of climate, crop, rangeland and livestock and the component models in capitals. Arrows indicate information flows.

tions to derive the potential pasture intake (see section 2.2.4). Finally, data on livestock and feed management, herd size and composition were used to set up LIVSIM for each cattle-owning household in the database.

2.2.2. Climate

Continuous daily climate data sets of 31 years were produced for a baseline and multiple futures following Ruane et al. (2015). Sets consisted of daily minimum and maximum temperatures, rainfall and solar radiation. The baseline (1980–2010) was built from the best available daily observations for Nkayi district, filled where necessary with the AgMERRA data set. For the future data sets, two representative concentration pathways (RCP), RCP4.5 and RCP8.5, were used. The future climate data sets for the mid-century period (2040–2070) were generated by perturbing the baseline set until monthly mean and deviation distributions best fitted the distributions computed with 29 GCMs (Ruane et al., 2015), representing the best part of the CMIP5 experiments at the time (Taylor et al., 2012).

The future projections consistently showed increases in temperature. Rainfall projections were variable across GCMs, but suggested a consistent decrease during the rain onset in October and November. In this study three GCMs were selected as representatives of a cluster of most detrimental scenarios (labelled "hot-dry"), least detrimental scenarios ("cool-wet") and median scenarios ("middle") (Ruane and McDermid, 2017). The "hot-dry" scenarios suggested daily mean temperature changes ranging from + 2 °C to + 2.8 °C and + 2.7 °C to + 3.6 °C and daily rainfall changes ranging from − 2% to − 20%, and − 5% to − 30% under RCP4.5 and RCP8.5 respectively. In the "cool-wet" scenarios temperature changed by + 1 °C to + 2 °C and + 1.5 °C to + 2.7 °C under RCP4.5 and RCP8.5, and daily rainfall changed by − 2% to + 12% under RCP4.5, and − 5% to + 22% under RCP8.5. The "middle" scenarios predicted daily mean temperature changes of + 1.7 °C to + 2.2 °C and + 2.5 °C to + 3 °C, as well as daily rainfall changes of − 7% to + 3% and − 10% to + 1% under RCP4.5 and RCP8.5 respectively.

2.2.3. Crop production

Grain and stover yields of maize, sorghum, groundnut and the fodder crop mucuna (*Mucuna pruriens* (L.) DC.) were simulated using the Agricultural Production Systems Simulator model (APSIM; Holzworth et al., 2014). Crop and soil data from previous experiments in the same region were used to calibrate and evaluate the model (Masikati et al., 2013). APSIM was run for all crops, all households and for each 31-year climate scenario. Household-specific settings were used for soil type, fertilizer rate and sowing window. Three common soil types in the area, further referred to as good, medium and poor soils, were inherently infertile Kalahari sands, differing in their organic carbon content and plant available water capacity (soil descriptions in Masikati (2011) and Masikati et al. (2015)).

2.2.4. Rangeland production and biomass availability

Grass production in the rangelands was simulated with the pasture module GRASP (McKeon et al., 2000; Rickert et al., 2000) implemented in APSIM. GRASP was used previously for assessing effects of climate change in a wide range of rangeland and pasture environments in Australia (McKeon et al., 2009). GRASP was run for the typical "medium" soil of the study region (Masikati, 2011) with a topsoil (0–15 cm) organic carbon content of 0.5% and plant available water content of 72 mm in the top one meter. The same 31-year climate data as for the crop simulations were used. The CO_2 fertilization effects were mimicked by adjusted GRASP parameters following Stokes et al. (2012). Model predictions were checked with literature values. As no detailed data on woody rangeland vegetation was available, we did not simulate browse biomass production, but estimated browse intake as explained below.

The monthly available grass biomass from the rangelands was calculated based on the simulated daily growth rate, averaged for every month. In our approach, the available biomass per animal depends strongly on the stocking rate, which was estimated at 2 ha TLU^{-1}, midway the reported range of 0.5 to 3.5 ha TLU^{-1} (Masikati, 2011; Homann-Kee Tui et al., 2013b), and not on rangeland

Table 1
Selected feed quality parameters in LIVSIM for the major feed resources in Nkayi district.
Sources: Masikati, 2011; SSA feeds database (https://vslp.org/ssafeed/, last visited 29 October 2016).

Feed resource	Parameter	Unit	Rainy season November–March	Early dry season April–June	Late dry season July–October
Grass	Dry matter content	g kg^{-1}	200	500	800
	Metabolizable energy	MJ kg^{-1}	10.3	8.7	6.5
	Crude protein	g kg^{-1}	134	100	40
	Dry matter digestibility	g g^{-1}	0.65	0.48	0.48
			Fresh residue season March–September	Old residue season October–February	
Maize stover	Dry matter content	g kg^{-1}	910	910	
	Metabolizable energy	MJ kg^{-1}	8.7	7.6	
	Crude protein	g kg^{-1}	45	40	
	Dry matter digestibility	g g^{-1}	0.58	0.5	
Sorghum stover	Dry matter content	g kg^{-1}	910	910	
	Metabolizable energy	MJ kg^{-1}	8.4	7.4	
	Crude protein	g kg^{-1}	63	56	
	Dry matter digestibility	g g^{-1}	0.63	0.55	
Groundnut haulms	Dry matter content	g kg^{-1}	920	920	
	Metabolizable energy	MJ kg^{-1}	8.9	8.9	
	Crude protein	g kg^{-1}	145	145	
	Dry matter digestibility	g g^{-1}	0.6	0.6	
Mucuna haulms	Dry matter content	g kg^{-1}	920	920	
	Metabolizable energy	MJ kg^{-1}	9.5	9.5	
	Crude protein	g kg^{-1}	150	150	
	Dry matter digestibility	g g^{-1}	0.65	0.65	

accessibility as in de Haan et al. (2016). A monthly loss rate of 20% was adopted to take account of senescence, trampling and a non-utilized fraction of the grass biomass. When available grass biomass provided less than the required dry matter (DM), 20% of the requirement was assumed to be met from browsing. This was based on reports from similar environments in Zimbabwe stating that cattle spend about 10 to 30% of their time browsing in the dry season (Scoones, 1995; Illius et al., 2000). The monthly DM requirement was derived from metabolizable energy (ME) requirements ranging from 45 to 65 MJ ME day^{-1} animal^{-1} and grass ME content ranging from 6.5 to 10.3 MJ kg^{-1} DM in the dry and the wet season respectively (Table 1).

2.2.5. Livestock production

In this study we focused on cattle and ignored other livestock species. Cattle production for each household in the database was simulated with the LIVSIM model (LIVestock SIMulator, Rufino et al., 2009), using the 31-year simulation results from the crop and rangeland models as input. Using a monthly time step, LIVSIM calculates the performance of every individual animal in the herd according to genetic potential, feed availability and quality, and herd management. Outputs comprise milk and manure production, body weight and herd dynamics, including offtake and mortality. LIVSIM was previously tested for Zimbabwean conditions with data from feeding trials with the Mashona breed (Rufino et al., 2011). For this study, LIVSIM was run for a local breed showing characteristics of the Nkone and Tuli breeds and their mixes. Breed characteristics, collected from literature (Garwe, 2001; Ngongoni et al., 2006; DAGRIS, 2007) and secondary data, were used for parameter derivation (Fig. 2, Table 2). Typical herd composition, herd and feeding management parameters were set based on expert knowledge and information derived from the household survey (Table 2). For example, in the model we specified that rangeland grazing takes place throughout the year, that crop residues are fed from June to December and that 80% of the stover is available for animal feeding. We further assumed that crop residues were fed exclusively to the animals of a particular farm.

Formal model validation requires data from feeding trials on feed intake in conjunction with for example data on milk production and body weight. For the local breed and feed resources in Nkayi no such data was available. Hence, "model sensibility" was tested by comparing
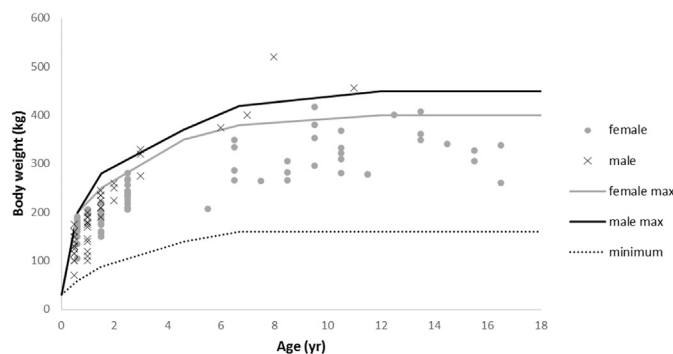
Fig. 2. Potential growth and minimum body weight curves for the local Nkone and Tuli breeds, as used in the LIVSIM model runs, with scatter points indicating measured body weight and age.
(Data source: Matopos research station)

Table 2
LIVSIM breed and livestock management parameters for the farming system in Nkayi district.
Sources: Garwe (2001); Ngongoni et al. (2006); DAGRIS (2007).

Parameter	Unit	Value
Maximum lactation length	months	10
Potential milk yield over the duration of one lactation	kg year^{-1}	linear interpolation between 200, 1100, 0 in month 0, 2 and 10 resp.
Milk fat content throughout lactation	g kg^{-1}	linear interpolation between 27, 32, 37, 40 in month 0, 3, 6, 10 resp.
Milk dry matter content	g kg^{-1}	130
Milk energy content	MJ kg^{-1}	19.6
Milk crude protein content	%	3.2
Weaning age	year	1.4
Daily walking distance	km	6.0
Maximum number of lactations[a]	#	4.0
Maximum "dry" period[a]	year	3.0

[a] Before the animal is replaced.

model outputs with farmer-reported livestock production from the household survey. In addition, a sensitivity analysis was performed on major feed and animal parameters and input variables, including productivity, metabolizable energy and crude protein content of crop residues and grass, milk yield potential, length of lactation, minimum and maximum body weight, daily walking distance, maximum number of lactations and maximum "dry" period before culling. Each parameter was varied from -50% to $+50\%$ of its original value with steps of 10%, while keeping the other parameters constant. The model was run with one animal and three farm stocking densities (2, 4 and 8 TLU ha^{-1}). For each parameter and stocking density, a sensitivity index (SI) was calculated based on the lifetime milk production (M, kg cow^{-1}) of the cow:

$$SI = \frac{M_{+0.5} - M_{-0.5}}{M_0}$$

where $M_{+0.5}$, $M_{-0.5}$, and M_0 refer to the lifetime milk production for parameter values that deviate $+50\%$, -50% and 0% from the original value respectively.

2.2.6. Scenario analysis

The modelling framework was run for each combination of seven climate scenarios (baseline and six future climates) and two management scenarios (current and improved) to assess the sensitivity to climate change and effects of adaptation. The current system was based on crop and livestock management information derived from the survey. For the improved system, effects of altered management were evaluated by comparing two adaptation packages (a and b). "Package a" consisted of improved soil fertility management with mineral fertilizer (20 kg N ha^{-1}) and manure (1 t ha^{-1}) on both maize and sorghum. "Package b" combined the former with the inclusion of grain and fodder legumes in rotation with the cereals. A minimum area of maize was kept to secure food self-sufficiency and the rest of the maize area was converted equally into groundnut and mucuna, resulting in new farm-specific cropland allocations. Mucuna biomass was used both as fodder (70%) and as mulch (30%), contributing to soil fertility. For all scenarios, simulated values of milk, manure, herd size, mortality and offtake were assessed for all households. A simple analysis of costs and benefits, following the procedures of Homann-Kee Tui et al. (2015a) and using local prices at the time of the survey, allowed calculating annual net revenues from the simulated livestock outputs of milk, offtake, draft power, and manure.

3. Results

3.1. Household characteristics

From the household database comprising 160 households, we omitted the extremely poor that do not own cattle (43% of the farm population). Average herd sizes for the remaining 91 households were 5.4 and 13.6 TLU for the poor (38% of the population) and non-poor households (19% of the population) respectively (Table 3). Cropland area and stocking density were significantly larger for non-poor farmers, compared to poor farmers. Herd composition, cropland allocation to the major crops and the proportions of farmers growing maize, sorghum and groundnut were similar for both groups (Table 3). Within the farm types there was still considerable variation in farm structural characteristics. For example, the stocking density ranged from 0.2 to 18.2 TLU ha^{-1} and from 1.7 to 38 TLU ha^{-1} in the poor and non-poor group respectively. As the farm stocking density determines fodder availability, thereby influencing livestock productivity, further analyses were differentiated by stocking density class (Table 3).

3.2. Model evaluation

The performance of the APSIM model was evaluated elsewhere

Table 3
Crop and livestock herd characteristics (averages with standard deviation in brackets) of poor and non-poor farm households in the Nkayi district.

	Poor	Non-poor
Number of households	61	30
Cropland area (ha)	2.0 (1.3)	2.7 (1.7)
% farmers growing maize	100	100
Cropland allocation to maize (%)	77 (22)	72 (23)
% farmers growing sorghum	33	43
Cropland allocation to sorghum (%)	9 (16)	14 (22)
% farmers growing groundnut	49	47
Cropland allocation to groundnut (%)	10 (13)	11 (15)
Total cattle TLU	5.4 (2.5)	13.6 (4.0)
Number of adult female cattle	1.7 (1.0)	4.5 (2.0)
Number of adult male cattle	1.1 (1.1)	2.1 (1.5)
Number of heifers and young steers	1.1 (1.1)	3.4 (1.7)
Number of calves	0.7 (0.9)	2.2 (1.7)
Stocking rate (TLU ha^{-1})	3.7 (3.0)	8.5 (8.3)

Stocking density (SD) classes	Number of households	Number of households
SD2: SD < 2.5 TLU ha^{-1}	24	4
SD4: SD ≥ 2.5 and SD < 5 TLU ha^{-1}	23	8
SD8: SD ≥ 5 TLU ha^{-1}	14	18

(Masikati et al., 2013, 2015). Average simulated grass production in the rangelands was 1358 kg ha^{-1} (ranging from 542 to 2199 kg ha^{-1}) for the baseline climate. The simulations corresponded with general productivity values for savanna vegetation (Rutherford, 1978) and with specific data from areas similar to the study site. For example, depending on the utilization rate, mean annual production for semi-arid bush savanna in South-East Zimbabwe ranged from 1217 to 2084 kg ha^{-1} for sites with annual rainfall of 520–620 mm (Kelly and Walker, 1976). Depending on soil texture and rainfall, grass production varied from 150–500 kg ha^{-1} to 700–2166 kg ha^{-1} (Scoones, 1995), in line with values of 683 to 2194 kg ha^{-1} for Natural Region IV of Zimbabwe (Dye and Spear, 1982) and with 1046 kg ha^{-1} for nearby Matopos (Masikati, 2011). We found a slope of 2.2 for the regression ($R^2 = 0.66$, p < 0.001) between seasonal rainfall and simulated biomass production, which is within the range of slopes of 1.3 for a less fertile Matopos Sandveld and 3.6 for a more fertile Matopos Thornveld (Dye and Spear, 1982), indicating that the effect of rainfall on grass production was adequately captured by the model.

Annual simulated milk production per farm (923 kg yr^{-1} on average across all households) was larger than the reported values (584 kg yr^{-1} on average) (Fig. 3a). Whereas the variability for the simulated data (CV = 0.48) was lower than for the observed data (CV = 0.89), the model captured the trend in the observed data with an R^2 of 0.53 (Fig. 3a). The simulation also overestimated the reported number of calves with an average calving rate of 0.36 against 0.20 for the reported values. Nevertheless, the R^2 value of 0.63 showed that the model captured a large part of the variability in the reported numbers (Fig. 3b). The average simulated cattle mortality rate (3%) was smaller than the average reported rate of 14%.

Although LIVSIM overestimated livestock production compared to the farmer-reported values, several factors call for a cautious interpretation. Firstly, the survey data were based on farmer recall, which is dubious as farmers typically do not measure animal production, use standard measuring devices, or keep records. Secondly, smallholders are often reluctant to disclose information about their animals, which are a wealth indicator. Thirdly, the LIVSIM model does not take into account animal losses due to disease, theft, or other stochastic factors. Nevertheless, LIVSIM captured the overall trend in reported values and the variability between households reasonably well.

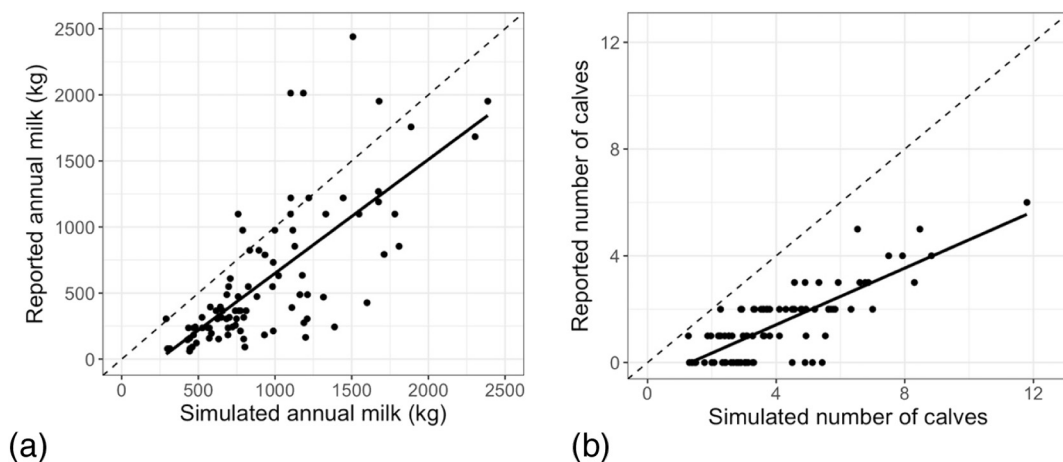The sensitivity analysis revealed that livestock productivity was less

Fig. 3. Scatterplots of reported versus simulated annual milk production (a) and number of calves born in the herd (b) for the current system and baseline climate. The dotted line is the 1:1 line and the full line is the regression line.

affected by changes in crop residue than in pasture parameters (Fig. 4), because of the larger contribution of pasture to the annual diet (Fig. 5). Similarly, pasture productivity was more determining for farms with larger stocking density. With respect to the breed-related parameters, the model was sensitive mostly to the potential milk yield (Fig. 4). Minimum and maximum body weight also influenced lifetime milk production, which was largest between − 40% and + 10% of the original parameter values. Management parameters were generally less important. Overall, LIVSIM is reacting to changes in parameter and input variables as expected and the model is relatively robust when parameter values stay within a reasonable range of uncertainty.

3.3. The livestock feedbase under baseline and future climate conditions

During the rainy and early dry seasons, cattle depended on range-land grazing, which was supplemented with on-farm fodder resources

from June to December (Fig. 5). Crop residues provided a varying proportion of the livestock diet, decreasing from 40% to 10% with increasing stocking density. Cattle experienced feed gaps from August up to October, especially in terms of energy and protein supply and particularly on densely stocked farms (Fig. 5).

Climate change is expected to affect both the rangeland and the on-farm fodder production (Fig. 1). Average grass production in the rangelands decreased only in the "hot-dry" scenarios, particularly under RCP8.5 (Fig. 6). In the "middle" scenario grass production did not deviate much from the baseline production, whereas in the "cool-wet" scenario, grass production improved. The year-to-year variability in grass production increased in all future climate scenarios: compared to the baseline (coefficient of variation (CV) 0.38), CVs increased by 20% in the "cool-wet" to 47% in the "hot-dry" scenarios of RCP8.5.

Average simulated stover yields ranged from 1.3 to 3.6 t ha^{-1} for maize and 1.4 to 1.9 t ha^{-1} for groundnut on poor to good soils
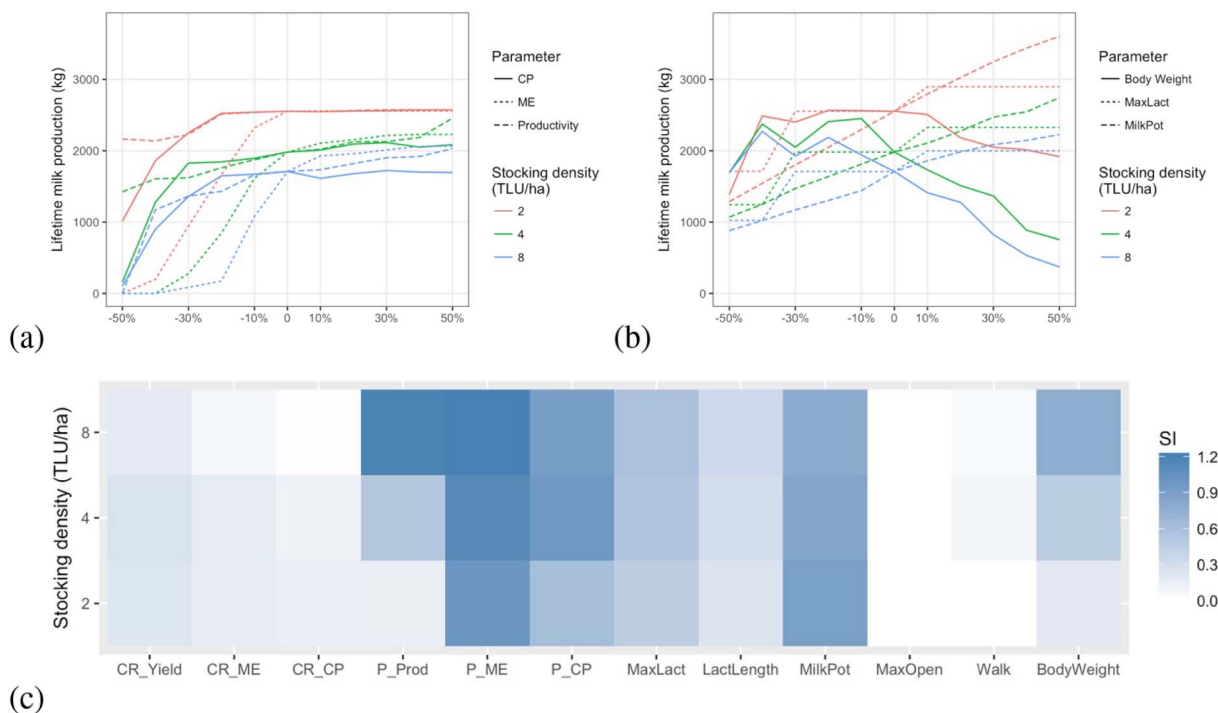

Fig. 4. Results of the sensitivity analysis with changes in lifetime milk production resulting from varying parameters (% deviation from the original value) related to pasture (a), breed and management parameters (b), and a heat map of the sensitivity index (SI) for twelve parameters and three stocking densities. Parameter abbreviations stand for CR: crop residue, P: pasture, ME: metabolizable energy, CP: crude protein, Prod: productivity, MaxLact: maximum number of lactations, LactLenght: length of the lactation, MilkPot: potential milk yield, MaxOpen: maximum period open, Walk: daily walking distance, BodyWeight: minimum and maximum body weight.
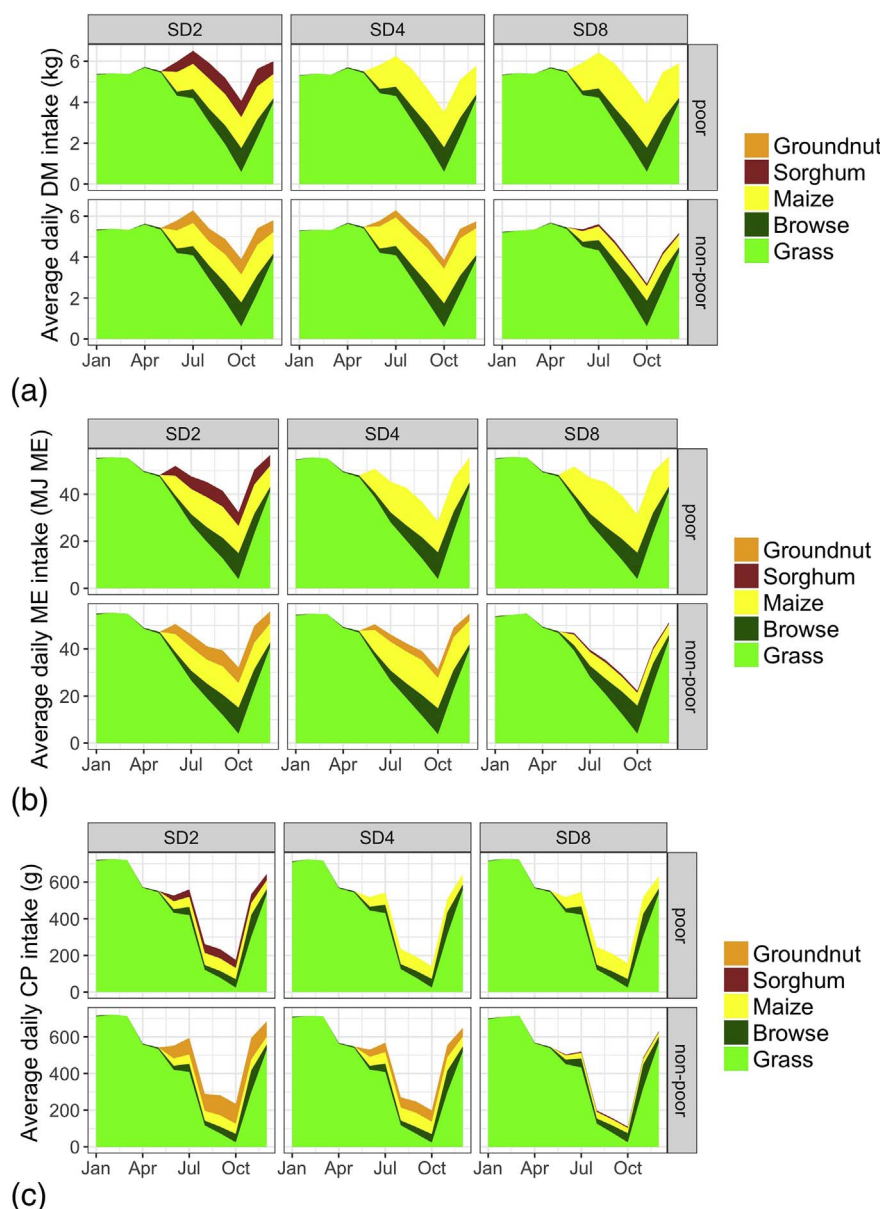
Fig. 5. Average daily dry matter (a), metabolizable energy (ME) (b) and crude protein (CP) (c) intake of different feed sources per animal for poor and non-poor representative households in three stocking density (SD) classes (see Table 3 for a description of farm types) for the current system and baseline climate. Differences in the contribution of the different crops depend on the farm-specific cropland allocation. Crop names refer to crop residues.

respectively in the baseline climate (Table 4). Sorghum was only grown on poor soils and stover yielded 1.5 t ha⁻¹. Climate change lowered average cereal stover yields by up to 5% and 12% for RCP4.5 and RCP8.5 respectively and most strongly in the "hot-dry" scenarios (Table 4). The year-to-year variability increased in the "hot-dry" scenarios, especially under RCP8.5, and decreased in the "cool-wet" scenarios. Average groundnut stover yield and its variability increased with climate change and more strongly with higher CO_2 concentration, except for the "hot-dry" scenario (Table 4). As a result of changes in feed production, the average annual fodder intake per animal decreased in the "hot-dry" scenarios by up to 9% compared to the baseline, with a stronger decrease on farms with higher stocking density. Conversely, in the "cool-wet" scenario of RCP8.5 an increase in fodder intake of up to 5% was simulated.

3.4. Livestock production under baseline and future climate conditions

Simulated body weight varied strongly throughout the year, with

cows losing weight in the dry season especially because of lack of protein in the feed, and gaining weight again in the wet season from plenty high quality fodder in the rangelands (Figs. 5 and 7, Table 1). Body weight fluctuations are a major cause of inefficiency, because the feed resources used to regain animal body condition after the dry season can not be used for production. Besides fluctuations within the year, livestock production also strongly varied between years. Coefficients of variation (CV) ranged between 0.4 and 0.6 for annual milk production, and 0.3 and 0.4 for net revenue in the baseline climate (Table 5), illustrating the riskiness of the livestock component in semi-arid mixed systems. Outputs linked to animal numbers, such as manure, varied less between years (Table 5). Livestock production further varied with herd size, which positively influenced livestock outputs in absolute terms, and with stocking density, which adversely affected individual animal productivity (Table 5). Net revenue, aggregating all livestock outputs and services, was larger and less variable in non-poor households, because larger herds better maintained a steady flow of outputs. In the current system, simulated milk, offtake and draft power each
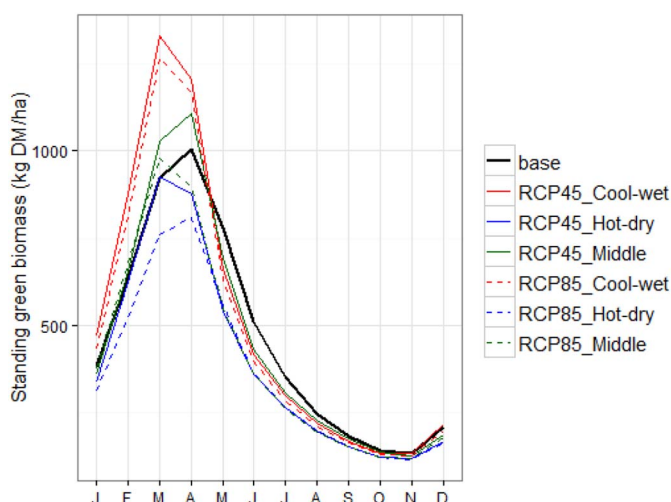
Fig. 6. Average simulated monthly grass biomass for different climate scenarios.
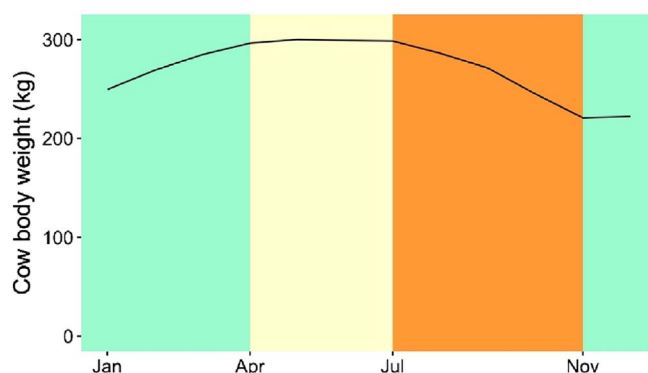

Fig. 7. Average simulated body weight evolution through the year for the current system and baseline climate. Colours indicate the seasons, with green the rainy season, yellow the early dry season and orange the late dry season. (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

contributed about one third of the annual revenue, whereas manure contributed very little because of its current very limited use. Because annual net revenues were based on simulated livestock output values, they overestimated farmers' reported income from cattle, which was 485 and 1363 US$ per year per farm for poor and non-poor farms respectively (Homann-Kee Tui et al., 2015a). Hence, in the following we pay more attention to relative changes than to absolute values.

Climate change had variable effects on simulated livestock production depending on the climate scenario and the farm characteristics. Livestock net revenues were negatively impacted in the "hot-dry" scenarios for nearly all farms with an average decrease compared to the baseline of 8 to 32% in RCP4.5 and 11 to 43% in RCP8.5. Households with larger stocking densities faced larger impacts (Fig. 8). Also the "middle" scenario of RCP8.5 resulted in a clear negative impact with relative declines in net revenue of 5–24%. Both "cool-wet" scenarios and the "middle" scenario of RCP4.5 resulted in overall positive, but small changes. However, even in these most positive scenarios about 25–30% of the farm population faced negative impacts. The "hot-dry" scenarios most strongly increased the risk farmers are facing, with a relative increase in the CV of up to 143% for RCP8.5. Again, this increase in risk is more pronounced for farms with larger stocking density. Also for the individual livestock outputs, the most severe impacts were simulated with the "hot-dry" scenario, which also led to the largest risk of losing animals (Fig. 9). The other scenarios resulted in milder effects, but due to the heterogeneity in the farm population, considerable proportions of the farmers still faced negative impacts.

Livestock functions depending on animal productivity (e.g. milk production, offtake) were more strongly affected by climate change than functions depending primarily on animal numbers (e.g. manure, draft power). This was because a decrease in fodder availability

immediately affected the energy and protein available for producing milk, gestation, and maintaining animal body condition. In contrast, the ability of animals to lose and re-gain weight cushioned the herd against changes in fodder availability. Across all livestock production indicators, stronger effects of climate change were simulated for the more densely stocked farms. Here, animals faced severe feed gaps in the dry season (Fig. 5), and any further drop in fodder production, further impaired their productivity, lowered their body condition and eventually increased mortality (Fig. 9).

3.5. Effects of the adaptation packages

The packages of intervention options improved the on-farm fodder production and quality. Compared to the current system, maize stover yields improved with "package a" by 27, 62 and 111% on good, medium and poor soils respectively because of organic and mineral fertilizer application. Thanks to additional rotational benefits in "package b", maize stover yields on good, medium and poor soils improved by 44, 112 and 201% compared to the baseline. As part of "package b", mucuna produced on average 4.3 to 5.0 t DM biomass per ha, depending on the soil type. Dry season feed gaps were weakly narrowed with "package a" and largely alleviated with "package b". Thanks to increased cereal stover production in "package a", total annual feed intake increased with 5% compared to the current system (Fig. 10a). However, as cereal stover is of poor quality, energy and protein gaps in the dry season remained. With "package b", leguminous stover from groundnut and mucuna improved not only the dry matter intake, but also the intake of metabolizable energy by 7–14% and crude protein by 10–26% (Fig. 10b), depending on the farm type. Only the most densely stocked farms still experienced a drop in crude protein intake in the dry season (results not shown).

Improvements in the feedbase because of the adaptation package were translated in increased livestock production and net revenues.

Table 4
Simulated average stover yield of maize, sorghum and groundnut for three representative soil types in the baseline climate with the coefficient of variation in brackets. For the future climate scenarios, average percentage change relative to the baseline yield and to the baseline coefficient of variation in brackets.

| Crop	Soil class	Baseline stover yield (kg ha^{-1})	RCP4.5			RCP8.5		
			Cool-wet	Hot-dry	Middle	Cool-wet	Hot-dry	Middle
Maize	Poor	1349 (0.08)	− 3 (− 20)	− 5 (3)	− 5 (0)	− 5 (− 17)	− 8 (20)	− 1 (0)
	Medium	2124 (0.08)	− 2 (− 2)	− 5 (18)	− 5 (3)	− 3 (− 4)	− 9 (50)	− 1 (− 3)
	Good	3611 (0.14)	0 (− 11)	− 5 (17)	− 4 (1)	− 1 (− 14)	− 12 (53)	− 2 (5)
Sorghum	Poor	1520 (0.10)	− 2 (− 2)	− 6 (1)	− 5 (0)	− 3 (− 5)	− 7 (12)	− 3 (1)
Groundnut	Poor	1408 (0.26)	29 (13)	21 (33)	29 (15)	36 (12)	15 (53)	32 (25)
	Medium	1951 (0.32)	24 (8)	17 (23)	24 (9)	32 (7)	10 (39)	28 (17)
	Good	1895 (0.28)	19 (6)	12 (24)	17 (9)	27 (6)	6 (41)	23 (20)

Table 5

Simulated average livestock production and productivity indicators for all poor and non-poor households in three stocking density (SD) classes (see Table 3 for a description of farm types). The standard deviation in brackets refers to the variation across households, whereas the coefficient of variation (C.V.) indicates the year-to-year variability. Offtake and mortality rate are calculated as the number of sold and dead animals in a year divided by the number of animals in the herd.

Indicator (unit)	Poor			Non-poor		
	SD2	SD4	SD8	SD2	SD4	SD8
Annual milk (kg farm^{-1})	644 (236)	691 (219)	785 (215)	1377 (231)	1677 (352)	1265 (409)
Annual milk C.V. (−)	0.48 (0.08)	0.53 (0.09)	0.54 (0.06)	0.42 (0.03)	0.46 (0.04)	0.57 (0.07)
Milk per lactation (kg cow^{-1})	329 (99)	289 (96)	281 (93)	318 (90)	328 (96)	263 (92)
Annual offtake (sold animals)	2.3 (1.0)	2.7 (0.9)	3.1 (0.8)	5.1 (0.7)	6.2 (1.0)	5.3 (1.9)
Annual offtake C.V. (−)	0.57 (0.13)	0.61 (0.15)	0.58 (0.10)	0.39 (0.04)	0.41 (0.08)	0.56 (0.11)
Offtake rate (−)	0.28 (0.05)	0.29 (0.04)	0.30 (0.04)	0.35 (0.02)	0.36 (0.03)	0.32 (0.04)
Annual mortality (animals)	0.16 (0.10)	0.28 (0.07)	0.38 (0.16)	0.32 (0.08)	0.34 (0.21)	0.66 (0.25)
Annual mortality C.V. (−)	3.02 (0.68)	2.34 (0.36)	2.46 (0.30)	2.68 (0.19)	3.14 (0.72)	2.42 (0.26)
Mortality rate (−)	0.022 (0.012)	0.036 (0.012)	0.040 (0.016)	0.023 (0.005)	0.021 (0.012)	0.044 (0.016)
Annual manure (kg DM farm^{-1})	6154 (1858)	6856 (1624)	7539 (1451)	11,719 (1799)	13,980 (2487)	11,969 (3426)
Annual manure C.V. (−)	0.24 (0.06)	0.23 (0.04)	0.22 (0.02)	0.16 (0.02)	0.17 (0.02)	0.22 (0.04)
Potential manure application rate (kg DM ha^{-1})	2919 (1033)	4674 (1189)	7445 (2238)	3253 (1428)	3935 (769)	9668 (5942)
Annual net revenue (US$)	1759 (652)	1989 (585)	2267 (544)	3725 (540)	4487 (785)	3755 (1197)
Annual net revenue C.V. (−)	0.36 (0.06)	0.38 (0.07)	0.37 (0.04)	0.29 (0.02)	0.30 (0.04)	0.39 (0.07)
Contribution to annual revenue (milk/offtake/draft/manure, %)	34/32/33/1	32/33/34/1	32/33/33/1	35/34/30/1	35/34/29/1	32/34/33/1

Compared to the current system, average annual milk production increased slightly with "package a" up to 6%, and more strongly up to 30% with "package b". Similarly the milk productivity of individual animals changed slightly with "package a", and more considerably by 15–30% with "package b" (Fig. 11). For offtake rates, smaller improvements were observed up to 15%. Aggregating all livestock outputs, net revenues increased by up to 12% and 20% compared to the current system with "package a" and "b" respectively (Fig. 8). In the improved system, year-to-year variation declined with the CV of net revenues up to 13% lower than in the current system. This decreased riskiness was also reflected in reduced chances of animal mortality (Fig. 9). The larger positive effect of "package b" compared to "package a" (Figs. 8 and 11) was attributed to the larger protein intake from leguminous stover, confirming that the current feed gaps are related more to feed quality than to feed quantity.

The negative effects of climate change on cereal crops were stronger in the improved system compared to the current system (results not shown). In the current system, soil nutrient limitations attenuated the potential adverse effects of temperature or drought stress, and this was less so in the improved system due to fertilizer and manure input. For livestock, the opposite was found with the improved system less sensitive to climate change, because the alleviation of feed gaps provided a buffer against drops in feed availability. For example in the "hot-dry" scenario the average decrease in net revenues amounted to 28% for the improved system instead of 43% for the current system. Whereas nearly all current farms faced declining net revenues with climate change, "package b" reduced this proportion to 60 and 76% of the farm population in RCP4.5 and RCP8.5.
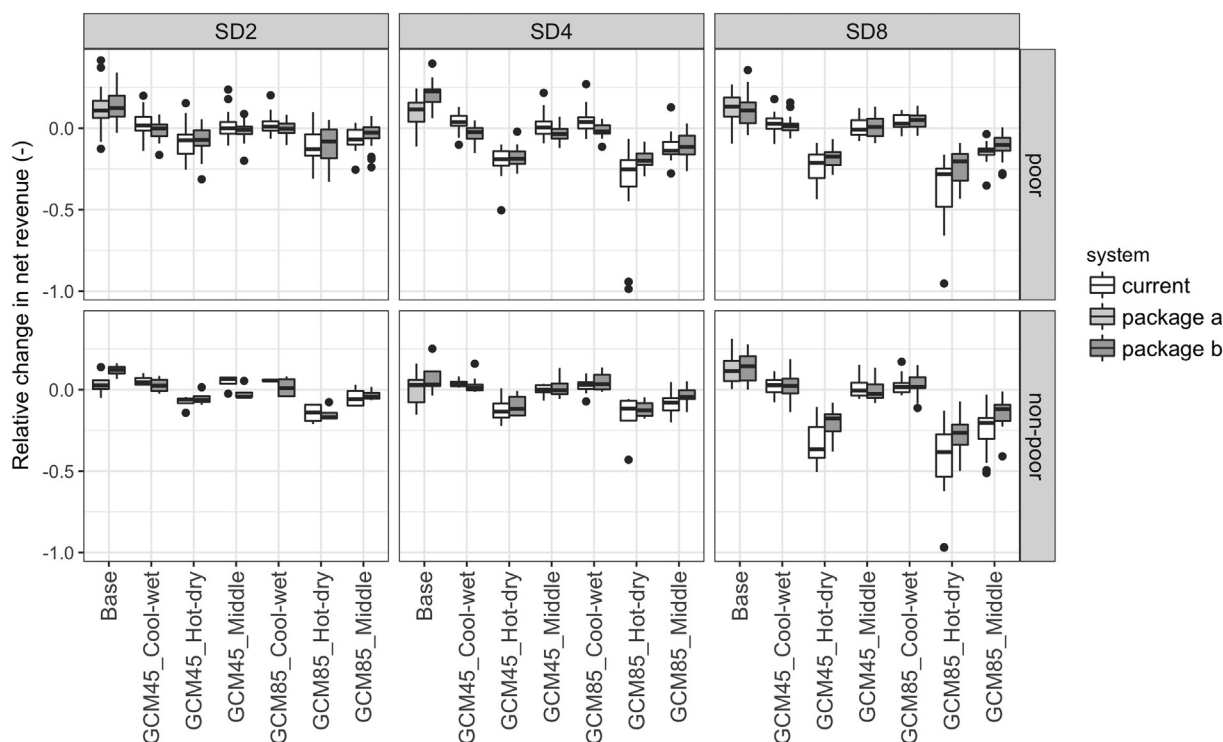
Fig. 8. Relative changes in simulated annual net revenue for all poor and non-poor households in three stocking density (SD) classes (see Table 3 for a description of farm types). Changes for the future climate scenarios (three typical climates for both the RCP4.5 and RCP8.5) are relative to the respective current and improved ("package b") system in the baseline climate. The leftmost boxplots indicate the relative change for "package a" and "package b" compared to the current system, both in the baseline climate.
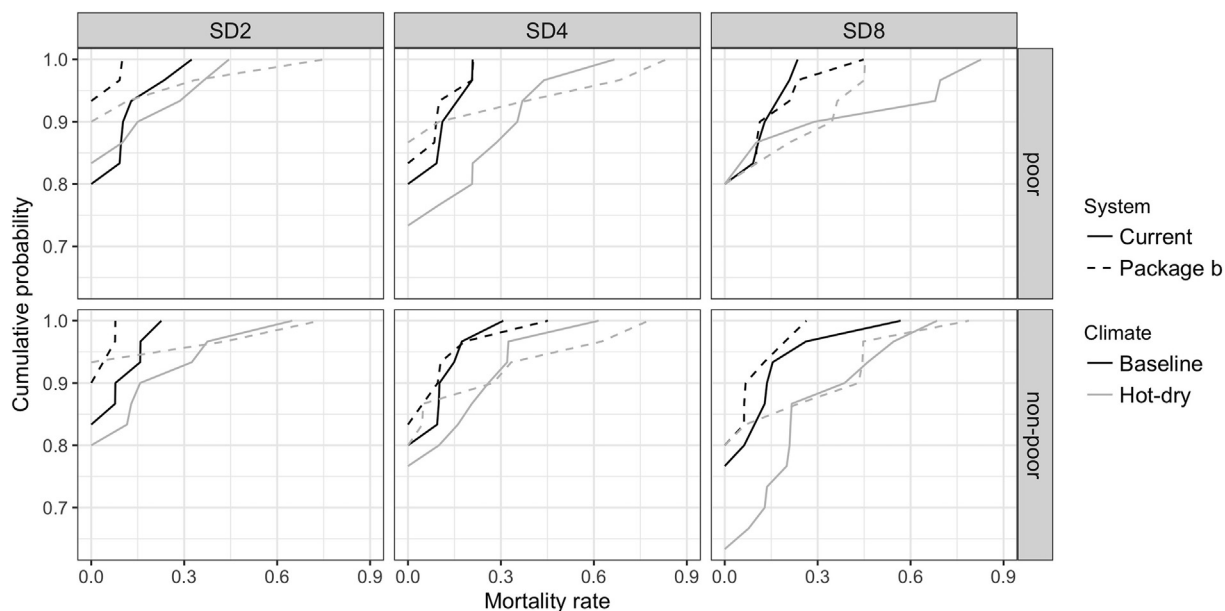
Fig. 9. Cumulative probability of mortality rates for poor and non-poor representative households in three stocking density (SD) classes (see Table 3 for a description of farm types) for the current and the improved system ("package b") in the baseline climate and the "hot-dry" climate of RCP8.5.

4. Discussion

Smallholder African farmers are widely believed to be vulnerable to the adverse effects of climate change (Descheemaeker et al., 2016a). More specifically, rainfed crop-livestock systems in arid and semiarid southern Africa were mapped as vulnerable hotspots (Thornton et al., 2008). Our integrated process-based modelling analysis revealed that the sensitivity to climate change of livestock contributes to this vulnerability in mixed systems. In several of the tested climate scenarios, livestock production and net revenues decreased, while at
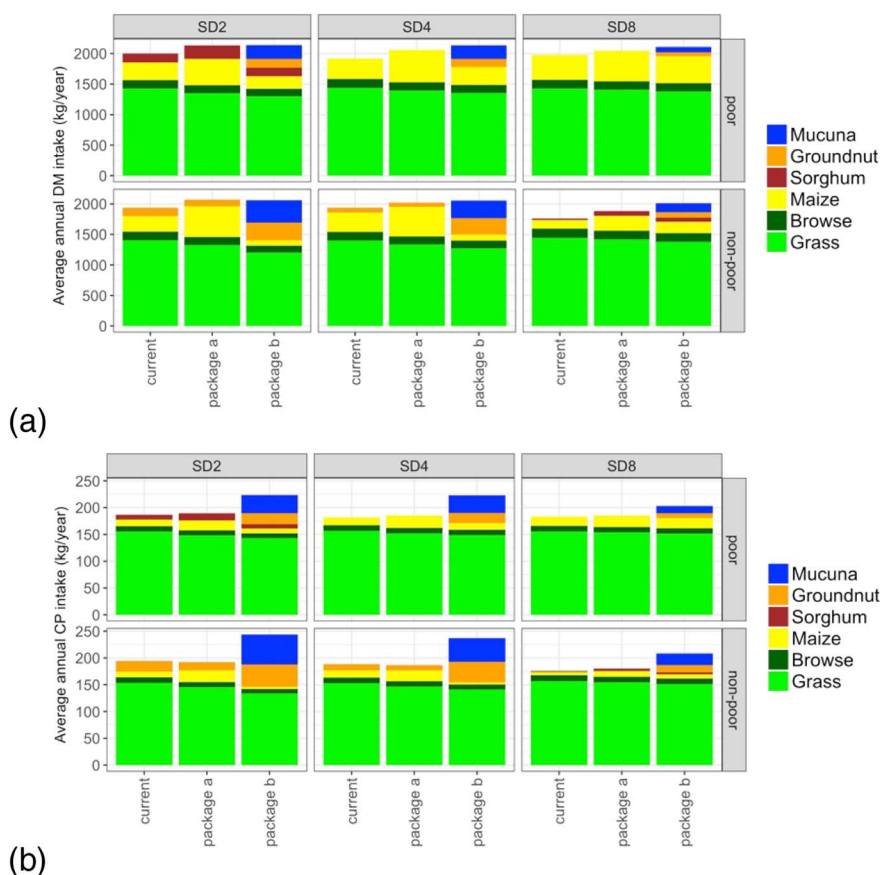

Fig. 10. Average annual dry matter (a), and crude protein (b) intake of different feed sources per animal for poor and non-poor representative households in three stocking density (SD) classes (see Table 3 for a description of farm types) for the current and the improved system with "package a" and "package b" in the baseline climate. Crop names refer to crop residues; proportions and changes in metabolizable energy not shown but comparable to dry matter in (a).
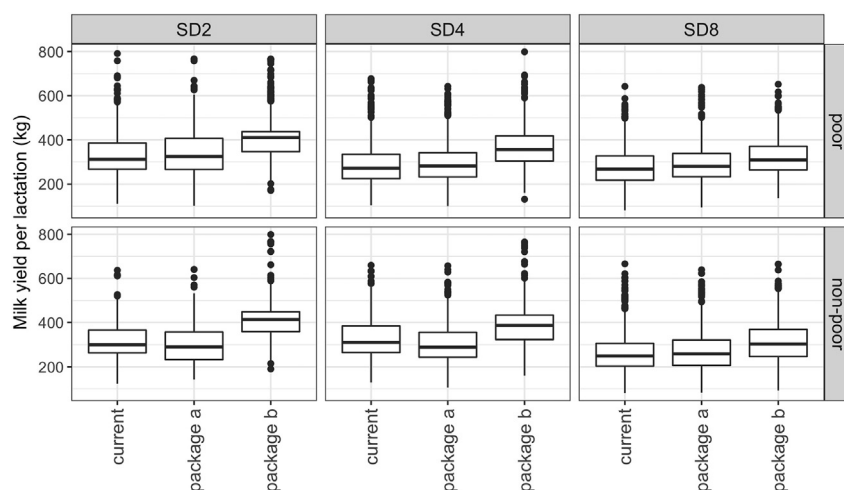
Fig. 11. Boxplots of milk yield per lactation for the current and the improved system with "package a" and "package b" in the baseline climate for all cows in typical poor and non-poor representative households in three stocking rate (SD) classes (see Table 3 for a description of farm types).

the same time risk increased. Like others highlighted for crops (e.g. Traore et al., 2017), we showed that also for livestock there is considerable uncertainty related to climate projections. Based on a range of potential changes, we quantified impacts from slightly positive in "cool-wet" to clearly negative in "hot-dry" scenarios. Varying income losses from African livestock farms were also predicted based on a Ricardian analysis (Seo and Mendelsohn, 2008). Using a similar approach for South Africa, Tibesigwa et al. (2015) found that specialized livestock and crop farmers were vulnerable to climate change, but crop-livestock farmers were not, a finding not supported by our analysis. Whereas some examples are available from global studies (e.g. Weindl et al., 2015) or other continents (e.g. Ghahramani and Moore, 2016), very few studies on African agriculture use dynamic simulation models to assess the likely impacts of climate change on livestock. An exception is the dynamic systems model of Dougill et al. (2010), predicting that climate change will exacerbate the decline in cattle herds and income from livestock in the Kalahari pastoral systems. Another, large-scale study on dryland livestock systems in West and East Africa used a combination of biophysical and socio-economic models showing that future drier conditions could reduce meat production by 14% (de Haan et al., 2016). Descriptions of climate change perceptions and current adaptations are more common in African studies, and several of those assess livestock systems (e.g. Silvestri et al., 2012; Megersa et al., 2014; Zampaligré et al., 2014). Although these studies do not quantitatively predict the future impact of climate change or adaptation, they offer insight into current trends and strategies. Farmers commonly perceive rainfall to become less in amount, less predictable, shorter in duration, and droughts to become more frequent, even though these perceptions are not always corroborated by climate records. Livestock keepers generally confirm the negative implications on feed availability and the repercussions on mortality, herd size, and animal performance (Silvestri et al., 2012; Megersa et al., 2014; Zampaligré et al., 2014). In other studies, negative impacts on livestock are inferred from anticipated changes in temperature, feed and water availability and animal diseases (e.g. Claessens et al., 2012). These and our findings on livestock systems' vulnerability alert preparing for a future in which livestock may become increasingly important because of widespread shifts from cropping to livestock keeping (Jones and Thornton, 2009) and increasing demands for livestock products (OECD/FAO, 2016; de Haan et al., 2016) across Africa.

We assessed the effects of climate change indirectly through the effects on the feedbase, an approach also followed in other process-based modelling studies (Weindl et al., 2015). This means firstly that our results depend on the representation of the feedbase. Simulated

intake from rangelands and crop residues contributed 78–91% and 22–9% of the annual diet respectively, depending on the stocking density. This is in the range reported for African mixed systems (Valbuena et al., 2015). However, the contributions of grazing and residues we found are larger and smaller respectively than the broad averages (50–55% for grazing and 30–35% for residues) reported by Herrero et al. (2013) for mixed rainfed systems in the African drylands. This discrepancy corroborates the relatively limited expansion of cultivated land, and the low livestock density and cropping intensity characterizing the agricultural system of Nkayi (Homann-Kee Tui et al., 2013b; Valbuena et al., 2015). Further, our diagnosis that the dry season feed gap is primarily caused by a lack of protein is confirmed in reports on the potential of legume forage options for Zimbabwean livestock production systems (Mapiye et al., 2007; Gwiriri et al., 2016).

Secondly, our indirect approach through the feedbase probably underestimated climate change impacts as the effects of heat stress, changes in water availability and pests and diseases (Thornton et al., 2009) were ignored. Another source of underestimation relates to the climate data, which did not capture possible changes in extreme events and variability. Although livestock can exploit variability in fodder availability to some extent, mortality and forced offtake in drought years lead to declines in herd size, which can persist for several seasons, unlike the shorter recovery time after crop failure. As climate change is expected to increase weather variability and frequency of extreme events (Porter et al., 2014), this risk needs to be better understood (Thornton et al., 2014), and investigated also for livestock (Naess and Bardsen, 2013). Further, coping with climate risk can be improved by including small ruminants, who play an important role for smallholders currently (Dube et al., 2014), and likely more so with climate change (Seo and Mendelsohn, 2008). As LIVSIM was recently adapted for goats (Amole et al., 2016), this can now be tested with the current modelling framework. Finally, instead of the grassland model used here, more detailed rangeland models (see Tietjen and Jeltsch (2007) for an overview) could be used if the necessary data on these environments becomes available for calibration and testing. Future developments could contribute by modelling browse intake, animal-vegetation interactions, and effects of climate change on woody vegetation production and species composition, and by accounting for spatial differentiation in productivity.

As we assessed impacts on livestock through effects on the feedbase, we tested feed interventions, acknowledging that adaptation may be achieved also through adapted breeds and livestock species, and animal and herd management (Silvestri et al., 2012). The adaptation packages consisted of combinations of technical options benefiting the three pillars of climate-smart agriculture through intensification and diversi-

fication (Thornton et al., 2007; Descheemaeker et al., 2016a). The packages were deemed feasible and promising in the current context of semi-arid Zimbabwe for the following reasons. No trade-off with food self-sufficiency was anticipated as a minimum area of cereals was maintained and 40% of the income from livestock is reported to be invested in food (Dube et al., 2014). The modest fertilizer rates (20 kg N ha^{-1}) were affordable and sufficient manure was produced to apply at least 1 t ha^{-1} on the cereal crops, if manure would be properly collected and stored (Table 5). Increasing the area of groundnut was a promising strategy as groundnut is becoming a cash crop in Zimbabwe (Homann-Kee Tui et al., 2015b), with also the haulms fetching high prices on African markets because of their feeding quality (Ayantunde et al., 2014). Dedicating part of the farm to the forage crop mucuna is gradually becoming a common practice in the study area (Homann-Kee Tui et al., 2015a), as livestock markets are fast developing (Orr and Mausch, 2014). Finally, leaving sufficient amounts of crop residues as a mulch for conservation agriculture instead of livestock feeding was deemed unfeasible in this region (Valbuena et al., 2015; Homann-Kee Tui et al., 2015a).

Our simulations suggested benefits from crop-livestock integration under climate risk. Integrated soil fertility management and cropland diversification with grain and forage legumes improved livestock productivity and reduced risk and sensitivity to climate change, corresponding to other reported expectations (Thornton and Herrero, 2014). Thanks to improved animal productivity (Fig. 11), current production could be maintained while reducing herd sizes by about 20%, illustrating that the tested intensification options can enable a shift towards keeping less, but more productive animals. This would have additional benefits in terms of reducing greenhouse gas emissions and water use per unit of product (Tarawali et al., 2011; Oosting et al., 2014), and reducing pressure on grazing resources. While alleviating feed gaps is effective and profitable in theory, it is complicated in practice, as illustrated by the widespread non-adoption of well-researched forage and feed solutions (Sumberg, 2002). Limits and constraints to the adoption of climate adaptation options were reviewed earlier for African crop-livestock systems in general (Descheemaeker et al., 2016a). For semi-arid Zimbabwe specifically, poor access to inputs, services and knowledge as well as poor access to output markets are major barriers (Homann-Kee Tui et al., 2015a). Furthermore, our study illustrated that the multi-functionality of livestock acts as an incentive for keeping large herds, which guarantee large net revenues. However, as larger stocking densities result in more severe feed gaps, they also impair animal productivity and increase the sensitivity to climate change. Nevertheless, stocking densities are unlikely to be reduced as long as open-access grazing policies prevail (Gebremedhin et al., 2004) and farmers have poor access to banking, insurance and farm mechanization services, which could replace the traditional livestock functions (Descheemaeker et al., 2016a). As smallholders are currently mostly disconnected from market dynamics, the increasing regional demand in animal protein is projected to be largely met by imports (OECD/FAO, 2016). Reversing this by enabling smallholders to tap into the potential through sustainable and climate-smart intensification, will require policies directed at market functioning and access, smallholders' capacity in market participation, infrastructure improvements, affordable input prices and credit schemes, tailored information and extension services, including meteorological services and early warning systems (Tarawali et al., 2011; Silvestri et al., 2012).

5. Conclusions

Our modelling framework integrating climate, crop, rangeland and farm management information enabled the assessment of climate change and adaptation effects on livestock production. We quantified effects on a range of livestock production indicators, which is essential in the smallholder context where farmers keep livestock for many functions besides meat and milk. Furthermore, our analysis captured

the heterogeneity in the farm population, showing that households are affected differently by changes in the climate and by management interventions. Through such disaggregated analysis, our framework generates essential information for effective targeting of solutions towards climate resilience. Comparing the effects of contrasting climate scenarios illustrated the typical uncertainty in climate impact assessments and highlighted the vulnerability of the livestock component of mixed systems. The current poor livestock productivity in the study area was related to dry-season feed gaps, as animals rely on grazing and low-quality crop residues. These feed gaps also explain the sensitivity to climate change. A comparison of two adaptation packages revealed that improving the quality of the feed through crop diversification with legumes raised animal productivity and reduced vulnerability. Climate-smart livestock production would require keeping less but better fed and hence more productive animals. However, such a transition needs to be enabled by changes in the institutional context, including rural banking and insurance services, functioning markets, and improved access to agricultural inputs.

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